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## Full Length Article

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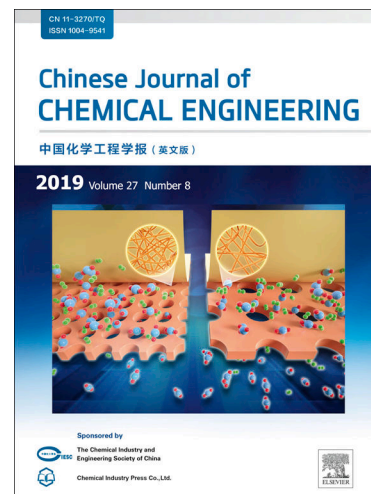
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**Insight into the promotional mechanism of Cu  
modification towards wide-temperature NH<sub>3</sub>-SCR  
performance of NbCe catalyst**

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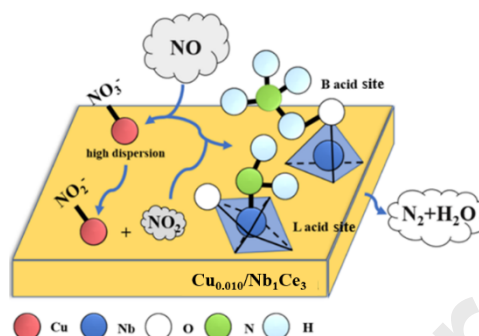
**Abstract**

A simple strategy of Cu modification was proposed to broaden the operation temperature window for NbCe catalyst. The best catalyst  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  presented over

90% NO conversion in a wide temperature range of 200–400 °C and exhibited an excellent H<sub>2</sub>O or/and SO<sub>2</sub> resistance at 275 °C. To understand the promotional mechanism of Cu modification, the correlation among the “activity-structure-property” were tried to establish systematically. Cu species highly dispersed on NbCe catalyst to serve as the active component. The strong interaction among Cu, Nb and Ce promoted the emergence of NbO<sub>4</sub> and induced more Brønsted acid sites. And Cu modification obviously enhanced the redox behavior of the NbCe catalyst. Besides, EPR probed the Cu species existed in the form of monomeric and dimeric Cu<sup>2+</sup>, the isolated Cu<sup>2+</sup> acted as catalytic active sites to promote the reaction:  $\text{Cu}^{2+}\text{-NO}_3^- + \text{NO}(\text{g}) \rightarrow \text{Cu}^{2+}\text{-NO}_2^- + \text{NO}_2(\text{g})$ . Then the generated NO<sub>2</sub> would accelerate the fast-SCR reaction process and thus facilitated the low-temperature deNO<sub>x</sub> efficiency. Moreover, surface nitrates became unstable and easy to decompose after Cu modification, thus providing additional adsorption and activation sites for NH<sub>3</sub>, and ensuring the improvement of catalytic activity at high temperature. Since the NH<sub>3</sub>-SCR reaction followed by E-R reaction pathway efficaciously over Cu<sub>0.010</sub>/Nb<sub>1</sub>Ce<sub>3</sub> catalyst, the excellent H<sub>2</sub>O and SO<sub>2</sub> resistance was as expected.

**Keywords:** NH<sub>3</sub>-SCR, NbCe catalyst, Cu modification, NO<sub>2</sub> promoting effect, fast-SCR, flue gas

## Graphical abstract



## 1. Introduction

The abatement of nitrogen oxides (NO<sub>x</sub>), one of the main contributors of atmospheric pollution, is still a challenging task worldwide [1-4]. Among the denitrification techniques developed in recent years, NH<sub>3</sub>-SCR has been adopted as the most promising and commercial employed procedure for NO<sub>x</sub> removal from stationary boilers and coal-fired plants [5, 6]. And the TiO<sub>2</sub>-anatase supported V<sub>2</sub>O<sub>5</sub>-WO<sub>3</sub> or V<sub>2</sub>O<sub>5</sub>-MoO<sub>3</sub> oxides have been commercialized for flue gas (300–400 °C) emitted from abovementioned stationary applications. Nevertheless, there are still some shortcomings, such as the biological toxicity of V<sub>2</sub>O<sub>5</sub> to human and eco-environment, narrow operation temperature window, and unselective oxidation of NH<sub>3</sub> to N<sub>2</sub>O, which constrain their application for nonelectric plant [7-9]. Therefore, it is compulsory to develop environmental-friendly NH<sub>3</sub>-SCR catalysts with exceptional low-temperature NO conversion, high N<sub>2</sub> selectivity and a broad operating temperature window as

alternatives.

Some transition metal oxides, such as  $\text{MnO}_x$ ,  $\text{CuO}$ ,  $\text{Fe}_2\text{O}_3$  and  $\text{CeO}_2$  have been found to be active for the SCR of  $\text{NO}$  with  $\text{NH}_3$  [10-13]. Among them,  $\text{CeO}_2$  has been extensively investigated due to (i) the excellent oxygen storage capacities (OSC) on the basis of the redox behavior between  $\text{Ce}^{3+}$  and  $\text{Ce}^{4+}$ ; (ii) the relatively easy formation of labile oxygen vacancies and the high mobility of bulk oxygen species [14]. Besides, the acid site and redox site are always highly valued in the  $\text{NH}_3$ -SCR process [15], so it's found that modifying  $\text{CeO}_2$ -based materials with acidic metal oxides ( $\text{WO}_3$ ,  $\text{Nb}_2\text{O}_5$  and  $\text{MoO}_3$  as examples) could dramatically promote their SCR performance [6, 16-18]. Among them, the  $\text{Nb}_2\text{O}_5$ - $\text{CeO}_2$  catalyst system was proven to show great potential for diesel  $\text{NO}_x$  emission control due to adequate activity and high thermal stability [19]. However, NbCe catalysts with desirable  $\text{NO}_x$  elimination performance at a broad operation temperature window are still in urgent need, and the poisoning tolerance to water vapor and sulfur oxides ubiquitously found in flue gas also waited to be enhanced [20]. The commonly reported Cu were widely used as active species, which exhibited superior low-temperature  $\text{NH}_3$ -SCR activity. As reported in the literature, the redox capacity of  $\text{CeO}_2$  could be promoted drastically by Cu modification [21-23]. And numerous studies manifested that the surface area, redox behavior and reactivity of  $\text{NH}_3$  together with absorbed nitrate were notably elevated for the binary Cu and Nb oxides [24-26].

In this work, Cu was adopted as active species to modify NbCe catalyst and the

optimal molar ratio of Cu/NbCe was explored. Then the promotional mechanism of Cu modification was comprehensively elucidated through clarifying the relationship among the “activity - structure - property” in depth.

## 2. Materials and Methods

### 2.1 Catalyst preparation

NbCe binary oxides were prepared by co-precipitation method. Desired amount of  $C_{10}H_5NbO_{20}$  and  $Ce(NO_3)_3 \cdot 6H_2O$  was adequately dissolved in deionized water, ammonia solution was then added dropwise to the mixture with continuous stirring until  $pH = 10$ . The mixture was further stirred for 3 h and then aged for 1 h. Afterwards, the precipitates were collected by filtration for several times and dried at  $100\text{ }^\circ\text{C}$  overnight and calcined at  $500\text{ }^\circ\text{C}$  for 4 h under air conditions. The catalysts were crushed and sieved to 40-60 meshes for activity tests. The molar ratios of Nb/Ce were 1/3, 1/1 and 3/1. The obtained catalysts were denoted as  $Nb_1Ce_3$ ,  $Nb_1Ce_1$  and  $Nb_3Ce_1$ .

Cu/NbCe ternary oxides were prepared by impregnation method. Desired amount of  $Cu(NO_3)_2 \cdot 6H_2O$  and the prepared NbCe powders were adequately dissolved in deionized water. The mixture was sufficiently stirred and heated at  $80\text{ }^\circ\text{C}$  until water was evaporated and a porous gel was formed. The gel was dried at  $100\text{ }^\circ\text{C}$  overnight and calcined at  $500\text{ }^\circ\text{C}$  for 4 h under air conditions. The catalysts were crushed and sieved to 40-60 meshes for activity tests. The obtained catalysts were denoted as  $Cu_{0.005}/Nb_1Ce_3$ ,  $Cu_{0.010}/Nb_1Ce_3$  and  $Cu_{0.030}/Nb_1Ce_3$  (molar ratio).

### 2.2 Characterization

X-ray diffraction (XRD) patterns were recorded on a Philips X'pert Pro diffractometer using Ni-filtered Cu K $\alpha$  radiation ( $\lambda = 0.15418$  nm) from 10° to 80°. The X-ray tube was operated at 40 kV and 30 mA.

X-ray fluorescence (XRF) measurements were performed on a ARL9800XP+ spectrometer to quantify the actual Cu contents of the samples.

N $_2$  adsorption-desorption measurements were performed at -196 °C using a Micromeritics ASAP-3020 analyzer. All samples were firstly degassed in vacuum at 300 °C. The specific surface areas of samples were calculated via Brunauer-Emmett-Teller (BET) method.

Laser Raman spectra were collected on a LabRAM Aramis (Japan Horiba) Laser Raman spectrometer with an Ar $^+$  laser beam with an emission line at 532 nm. The laser power was 10 mW.

X-ray Photoelectron Spectroscopy (XPS) experiments were performed on a PHI 5000 Versa Probe system equipped with a monochromatic Al X-ray radiation source (1486.6 eV) with an accelerating power of 15 kW. Binding energies (BE) of all the elements were calibrated with C1s at 284.6 eV.

H $_2$ -temperature-programed reduction (H $_2$ -TPR) was performed in a quartz U-tube reactor connected to a thermal conductivity detector (TCD). An H $_2$ -Ar (7% H $_2$ ) mixture was used as reductant. The 10 mg samples were pretreated with purified N $_2$  at 200 °C for 1 h. The range of measuring temperature was 20 °C to 800 °C at a rate of 10 °C·min $^{-1}$ .

Electron paramagnetic resonance (EPR) spectra were recorded by a Bruker EMX

spectrometer using 100 kHz modulation and 4-G standard modulation width.

Temperature programmed desorption (TPD) measurements were carried out in a fixed-bed quartz reactor equipped with a FTIR spectrometer to monitor outlet gas signal and the interval of spectrum collecting was set at 30 s. For NH<sub>3</sub>/NO<sub>2</sub>-TPD experiment: the 100 mg samples were exposed to 0.5‰ NH<sub>3</sub> or 0.5‰ NO<sub>2</sub> (gas mixture of 50 ml·min<sup>-1</sup>, Ar balanced) until saturation at ambient temperature. Subsequently, purged the samples with the flowing high purified Ar to remove gaseous and weakly absorbed species. Finally, the samples were heated to 500/400 °C at a rate of 10 °C·min<sup>-1</sup> to record the NH<sub>3</sub> or NO<sub>2</sub> signal.

*in situ* DRIFTS experiments were conducted on a Nicolet Nexus 5700 FTIR spectrometer. The spectra were collected by an MCT detector and presented as a Kubelka-Munk function. The resolution was 4 cm<sup>-1</sup> and 32 scans were taken. Before collecting the background spectra, the powder samples were pretreated with high purified N<sub>2</sub> at 400 °C. Background spectra were then collected in a flowing high purified N<sub>2</sub> stream as the powder samples cooled to each desired temperature. The sample spectra were then obtained after the subtraction of the background spectra at each desired temperature. The samples were exposed to a controlled stream of specific gas mixture for a certain amount of time: 3‰ NH<sub>3</sub>; 3‰ NO + 5 % (volume) O<sub>2</sub>. All of the reaction gasses were balanced with N<sub>2</sub> at a flow rate of 50 ml·min<sup>-1</sup>.

### 2.3 SCR performance measurements

The NH<sub>3</sub>-SCR activity and N<sub>2</sub> selectivity on the as-prepared samples were

measured in a fixed-bed quartz flow reactor with 200 mg of catalyst. The feed gases were composed of 0.5% NO, 0.5% NH<sub>3</sub>, 5 % (volume) O<sub>2</sub>, 0.2% SO<sub>2</sub> (when used), 5 % (volume) H<sub>2</sub>O (when used) and balanced with Ar. The total flow rate was 200 ml·min<sup>-1</sup> and the WHSV (weight hourly space velocity) was 60,000 ml·g<sup>-1</sup>·h<sup>-1</sup>. Before starting the activity test, the catalysts were pretreated in Ar at 200 °C for 1 h. The outlet gas concentrations were measured by an online Thermofisher IS10 FTIR spectrometer. The NO conversion and N<sub>2</sub> selectivity of the NH<sub>3</sub>-SCR reaction were determined according to the following equations:

$$\text{NO conversion} = \frac{[\text{NO}]_{\text{in}} - [\text{NO}]_{\text{out}}}{[\text{NO}]_{\text{in}}} \times 100\% \quad (1)$$

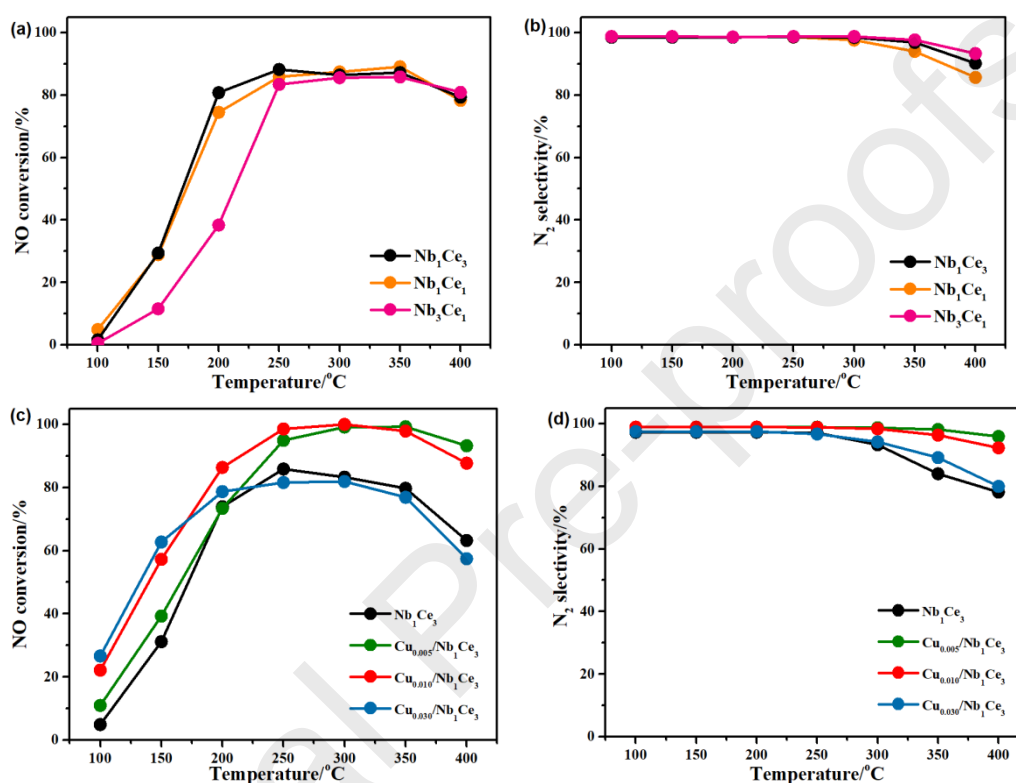
$$\text{N}_2 \text{ selectivity} = \frac{[\text{NO}]_{\text{in}} - [\text{NO}]_{\text{out}} + [\text{NH}_3]_{\text{in}} - [\text{NH}_3]_{\text{out}} - [\text{NO}_2]_{\text{out}} - 2[\text{N}_2\text{O}]_{\text{out}}}{[\text{NO}]_{\text{in}} - [\text{NO}]_{\text{out}} + [\text{NH}_3]_{\text{in}} - [\text{NH}_3]_{\text{out}}} \times 100\% \quad (2)$$

### 3. Results and Discussion

#### 3.1 Catalytic performance

The NO conversion and N<sub>2</sub> selectivity over Nb<sub>1</sub>Ce<sub>3</sub>, Nb<sub>1</sub>Ce<sub>1</sub> and Nb<sub>3</sub>Ce<sub>1</sub> were evaluated in broad temperature of 100 °C to 400 °C, and the corresponding results were shown in Fig. 1(a-b). All the NbCe catalysts had the favorable N<sub>2</sub> selectivity and the Nb<sub>1</sub>Ce<sub>3</sub> catalyst displayed the best NO removal efficiency in the range of 100 °C to 400 °C, so it was adopted as the support in the following study. Then the catalytic performance over Cu<sub>0.005</sub>/Nb<sub>1</sub>Ce<sub>3</sub>, Cu<sub>0.010</sub>/Nb<sub>1</sub>Ce<sub>3</sub> and Cu<sub>0.030</sub>/Nb<sub>1</sub>Ce<sub>3</sub> were reported in Fig. 1(c-d). It could be clearly seen that the Cu modification efficaciously enhanced the denitration efficiency and N<sub>2</sub> selectivity in a wide temperature. As the Cu adding amount increased, the NH<sub>3</sub>-SCR activity on Cu/Nb<sub>1</sub>Ce<sub>3</sub> catalysts initially increased and

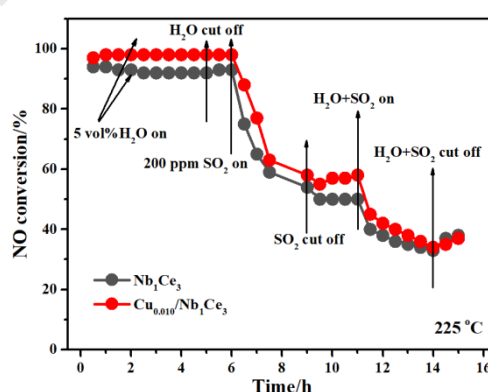
then decreased.  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  performed the supreme  $\text{NH}_3$ -SCR activity in the entire temperature range, over which exceeding 90% NO conversion was accomplished from 200 °C to 400 °C. It was concluded that the proper amount of Cu modification significantly boosted the DeNO<sub>x</sub> efficiency of  $\text{Nb}_1\text{Ce}_3$  catalyst.



**Fig. 1.** NO conversion and  $\text{N}_2$  selectivity of (a-b) NbCe and (c-d) Cu/Nb<sub>1</sub>Ce<sub>3</sub> as a function of the reaction temperature. (The feeding gas contained 0.5‰ of  $\text{NH}_3$ , 0.5‰ of NO and 5 % (volume)  $\text{O}_2$  balanced with Ar. Weight hourly space velocity (WHSV) = 60,000  $\text{ml}\cdot\text{g}^{-1}\cdot\text{h}^{-1}$ .)

In the practical  $\text{NH}_3$ -SCR reaction,  $\text{SO}_2$  and  $\text{H}_2\text{O}$  are considered as main deactivating species hence the excellent  $\text{SO}_2$  and  $\text{H}_2\text{O}$  durability of catalyst is highly required [27]. As illustrated in Fig. 2,  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  and  $\text{Nb}_1\text{Ce}_3$  were implemented to estimate the  $\text{SO}_2$  and/or  $\text{H}_2\text{O}$  resistance. The  $\text{SO}_2$  and  $\text{H}_2\text{O}$  durability test was conducted at 275 °C and the reaction was operated for 5 h before exposing to  $\text{H}_2\text{O}$ . The addition

of 5 % (volume)  $\text{H}_2\text{O}$  had a negligible inhibition effect on the SCR performance and the NO conversion restored to initial state once the  $\text{H}_2\text{O}$  supply was turned off, which indicated that the deactivation of  $\text{H}_2\text{O}$  was reversible. When 0.2%  $\text{SO}_2$  was introduced into the feed for 7 h, the activity of  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  only declined from 100% to 95% in first one hour and then stabilized. However, the activity of  $\text{Nb}_1\text{Ce}_3$  continued to drop from 85% to 75% until the  $\text{SO}_2$  was cut off. Then both  $\text{SO}_2$  and  $\text{H}_2\text{O}$  were introduced into the feed to examine the combined effect and the de $\text{NO}_x$  efficiency of  $\text{Nb}_1\text{Ce}_3$  exhibited sharply impaired. Its activity dropped from 80% to 45% within one hour, and then slowly recovered to 75% in the next 11 h. The corresponding NO conversion of  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  decreased to 90% after 12 h, and restored to almost 95% when the  $\text{SO}_2$  and  $\text{H}_2\text{O}$  supply was turned off, suggesting the deactivated  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  catalyst was easy to regenerate. Thus, it could be concluded that the  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  catalyst still displayed outstanding  $\text{NH}_3$ -SCR of NO performance even in the presence of  $\text{SO}_2$  and  $\text{H}_2\text{O}$ .

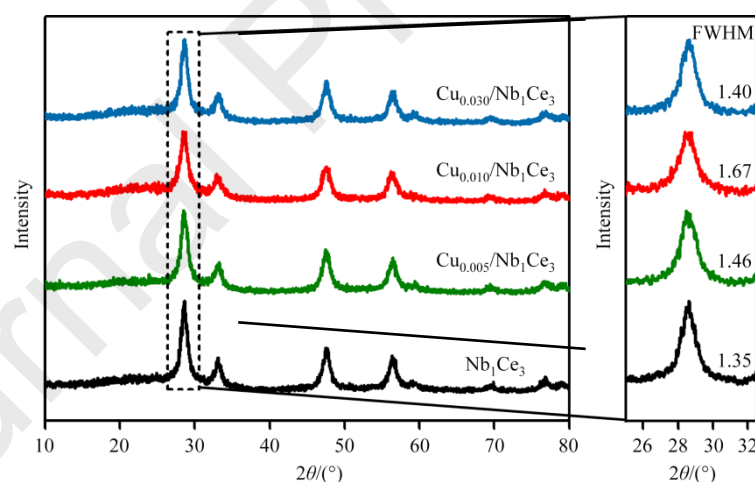


**Fig. 2.** Tolerance tests over  $\text{Nb}_1\text{Ce}_3$  and  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  as a function of the reaction time. (The feeding gas contained 0.5% of  $\text{NH}_3$ , 0.5% of  $\text{NO}$ , 5 % (volume)  $\text{O}_2$ , 5 % (volume)  $\text{H}_2\text{O}$  (when used) and 0.2% of  $\text{SO}_2$  (when used) balanced with Ar. Weight hourly space velocity (WHSV) =

60,000 ml·g<sup>-1</sup>·h<sup>-1</sup>.)

### 3.2 Physicochemical characterizations

The XRD patterns of the catalysts were presented in Fig. 3. All the catalysts provided typical diffraction peaks for the CeO<sub>2</sub> cubic phase (JCPDS 34-0394) and no diffraction peaks attributed to Nb species were detected, indicating that Nb species existed as amorphous. After Cu modification, the diffraction peaks of Cu/Nb<sub>1</sub>Ce<sub>3</sub> catalysts were broader than Nb<sub>1</sub>Ce<sub>3</sub> catalysts (clearer in the enlarged pattern attached to and FWHM obtained by fitting a Gaussian function) and no bands intensity ascribed to CuO were observed. This indicated that Cu ions inserted into the lattice of the Nb<sub>1</sub>Ce<sub>3</sub> catalysts or was highly dispersed on the surface of Cu/Nb<sub>1</sub>Ce<sub>3</sub> catalysts [28].



**Fig. 3.** XRD patterns of Nb<sub>1</sub>Ce<sub>3</sub> and Cu/Nb<sub>1</sub>Ce<sub>3</sub>.

The actual molar ratio of Cu to Nb<sub>1</sub>Ce<sub>3</sub> in the as prepared Cu/Nb<sub>1</sub>Ce<sub>3</sub> samples were detected by XRF analysis and the results were listed in Table 1. And they were basically consistent with the theoretical value, which proved that our samples were successfully prepared.

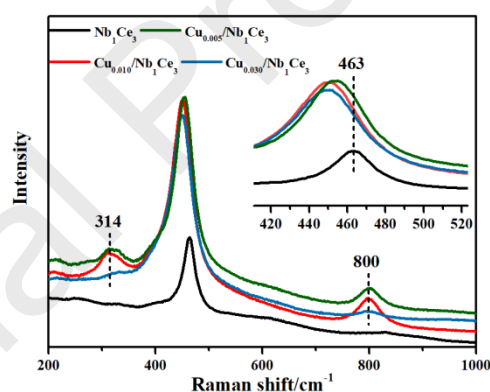
The BET surface area, pore volume, and average pore size derived from N<sub>2</sub> physisorption results were shown in Table 1. It was noticeably that the  $S_{\text{BET}}$  of Cu/Nb<sub>1</sub>Ce<sub>3</sub> catalysts gradually decreased with the increase of Cu addition yet its catalytic activities significantly preceded that of the Nb<sub>1</sub>Ce<sub>3</sub> catalyst. This indicated that the SCR performance did not directly depend on the specific surface area in such catalyst system. In addition, the pore volume of all samples decreased from 0.10 to 0.08 cm<sup>3</sup>·g<sup>-1</sup> while the pore size changed conversely with Cu modification. This manifested that the Cu species entered the pores of Nb<sub>1</sub>Ce<sub>3</sub> and caused the plugging to a certain extent.

**Table 1.** Physicochemical properties of Nb<sub>1</sub>Ce<sub>3</sub> and Cu/Nb<sub>1</sub>Ce<sub>3</sub>.

| Samples  | Cu<br>actual ratio | $S_{\text{BET}}/\text{m}^2\cdot\text{g}^{-1}$ | Pore volume/cm <sup>3</sup> ·g <sup>-1</sup> | Pore size/nm |
|--|--------------------|---|--|--------------|
| Nb <sub>1</sub> Ce <sub>3</sub>                      | -                  | 66.5  | 0.12   | 3.49         |
| Cu <sub>0.005</sub> /Nb <sub>1</sub> Ce <sub>3</sub> | 0.0067             | 49.3  | 0.10   | 2.88         |
| Cu <sub>0.010</sub> /Nb <sub>1</sub> Ce <sub>3</sub> | 0.013              | 35.1  | 0.09   | 3.31         |
| Cu <sub>0.030</sub> /Nb <sub>1</sub> Ce <sub>3</sub> | 0.033              | 28.4  | 0.08   | 3.41         |

Raman spectra were also collected to further explain the catalyst structure and depicted in Fig. 4. An intensive band centered at 463 cm<sup>-1</sup> was observed on Nb<sub>1</sub>Ce<sub>3</sub>, which was attributed to the F<sub>2g</sub> mode characteristic of the fluorite structure of ceria [29]. As for Cu/Nb<sub>1</sub>Ce<sub>3</sub>, the CuO bands were completely absent, illuminating the amorphous state of CuO species as the XRD patterns suggested. The band at 446 cm<sup>-1</sup> ascribed to

$F_{2g}$  showed an obvious shift toward lower wavenumber, signifying a decrease in the Ce-O bond order. This illustrated that the Cu-O-Ce solid solution may occur due to the Cu modification. In addition, two bands detected at 314 and 800  $\text{cm}^{-1}$  were ascribed to the  $\text{Nb}_2\text{O}_5$  phase [30], indicative of the existence of the Nb-O symmetric modes of the  $\text{NbO}_4$  tetrahedral structure.  $\text{NbO}_4$  was deemed as a highly active site, which was conducive to the adsorption and activation of  $\text{NH}_3$  [19]. It should be noted that the intensity of  $\text{NbO}_4$  tetrahedral coordination was the most obvious in  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  compared with other samples and this further corroborated the above activity test results. And the weakening of Ce-O bond and the appearance of  $\text{NbO}_4$  indicated that there was a strong interaction among Cu, Nb and Ce species.



**Fig. 4.** Raman spectra of  $\text{Nb}_1\text{Ce}_3$  and  $\text{Cu}/\text{Nb}_1\text{Ce}_3$ .

### 3.3 Surface elements properties

To better understand the surface chemical states of the catalysts, XPS measurements were conducted and the results were displayed in Fig. 5. Ce 3d, Nb 3d and O 1s were analyzed and the main surface metal content was summarized in Table 2.

**Table 2.** Surface atomic ratio of Nb<sub>1</sub>Ce<sub>3</sub> and Cu/Nb<sub>1</sub>Ce<sub>3</sub>.

| Samples  | Cu            | Surface atomic ratio/%                                  |   |   |
|--|---------------|---|---|---|
|  | atomic        | Ce <sup>3+</sup> /(Ce <sup>3+</sup> +Ce <sup>4+</sup> ) | O <sub>α</sub> /(O <sub>α</sub> +O <sub>β</sub> ) | Nb <sup>5+</sup> /(Nb <sup>5+</sup> +Nb <sup>4+</sup> ) |
|  | concentration |   |   |   |
|  | /(%)(mass)    |   |   |   |
| Nb <sub>1</sub> Ce <sub>3</sub>                      | -             | 26.4  | 37.3  | 73.7  |
| Cu <sub>0.005</sub> /Nb <sub>1</sub> Ce <sub>3</sub> | 1.36          | 25.9  | 27.6  | 76.6  |
| Cu <sub>0.010</sub> /Nb <sub>1</sub> Ce <sub>3</sub> | 2.20          | 27.8  | 33.1  | 83.6  |
| Cu <sub>0.030</sub> /Nb <sub>1</sub> Ce <sub>3</sub> | 3.28          | 27.6  | 33.7  | 71.5  |

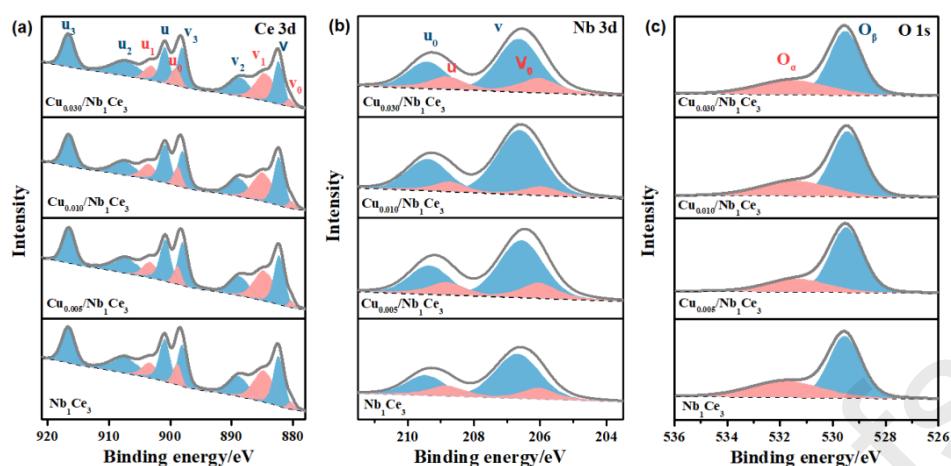
Due to the low Cu content, its peaks in XPS were greatly affected by the signal-to-noise ratio, so the atomic concentration was analyzed. As the Cu species addition increased, the concentration of Cu atoms on the surface of the catalyst gradually increased as well. However, it changed slowly (2.20% vs 3.28%) when the Cu loading exceeded 1 %(mass), which might be attributed to the occurrence of agglomerated Cu species. The agglomerated Cu species would lead to the oxidation of NH<sub>3</sub> [31], which explained the reason why the catalytic activity of the Cu<sub>0.030</sub>/Nb<sub>1</sub>Ce<sub>3</sub> catalyst degraded at high temperature.

The spectrum of Ce 3d was deconvoluted into ten peaks and the corresponding results were given in Fig. 5(a). The deconvoluted peaks labeled as “u” symbolized the Ce 3d<sub>3/2</sub> spin-orbit components, and those donated as “v” symbolized the Ce 3d<sub>5/2</sub> spin-orbit components. And the peak labeled as v<sub>0</sub>, v', u<sub>0</sub>, and u' were assigned to the Ce<sup>3+</sup> species, peaks donated as v, v'', v''', u, u'', and u''' belonged to the Ce<sup>4+</sup> species

[26].  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  acquired the highest  $\text{Ce}^{3+}/(\text{Ce}^{4+}+\text{Ce}^{3+})$  ratio compared with  $\text{Nb}_1\text{Ce}_3$  and other Cu modified catalysts. A higher concentration of surface  $\text{Ce}^{3+}$  might be beneficial for the enhanced  $\text{NH}_3$ -SCR performance. It demonstrated that the addition of Cu effectively transformed the  $\text{Ce}^{4+}$  to  $\text{Ce}^{3+}$  on the surface of  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  through the interaction between Cu and Ce species [22].

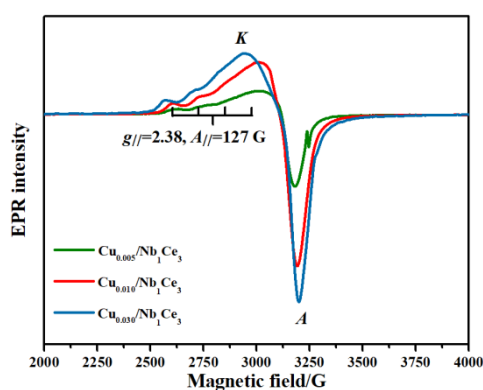
Nb 3d spectra were presented in Fig. 5(b) and the binding energy of Nb  $3d_{5/2}$  located at 206.8 and 205.6 eV could be ascribed to the  $\text{Nb}^{5+}$  and  $\text{Nb}^{4+}$  species, indicating that Nb species on  $\text{Nb}_1\text{Ce}_3$  and  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  were mainly in the highest valence state [32].  $\text{Nb}^{5+}$  was easier to enter the tetrahedral vacancy of oxygen due to its smaller radius than  $\text{Nb}^{4+}$  and thus the active  $\text{NbO}_4$  species appeared. The proper content of Cu modification facilitated  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  to acquire the most  $\text{Nb}^{5+}/(\text{Nb}^{5+}+\text{Nb}^{4+})$  ratio through the interaction between Cu and Nb.

As depicted in Fig. 5(c), the XPS spectra of O 1s were fitted into two peaks. The O 1s binding energy located at 530.4-531.1 eV was attributed to the chemisorbed oxygen species including  $\text{O}^-$  and  $\text{O}_2^-$  (donated as  $\text{O}_\alpha$ ) and the binding energy at 529.4-529.9 eV was ascribed to the lattice oxygen species (denoted as  $\text{O}_\beta$ ) [33].  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  acquired the highest  $\text{O}_\alpha/(\text{O}_\alpha+\text{O}_\beta)$  ratio, demonstrating that the  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  catalyst possessed a large amount of the surface chemisorbed oxygen species, which was favor of NO oxidation to  $\text{NO}_2$  [24]. Thus, the abundant amount of chemisorbed oxygen species on the surface of  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  could contribute to the fast-SCR reaction and consequently much better activity at low temperature.



**Fig. 5.** XPS spectra of (a) Ce 3d, (b) Nb 3d and (c) O 1s for  $\text{Nb}_1\text{Ce}_3$  and  $\text{Cu}/\text{Nb}_1\text{Ce}_3$ .

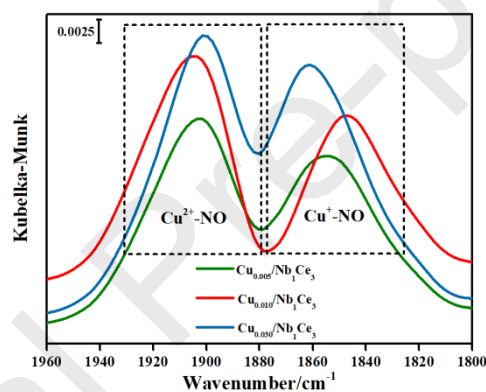
A deeper understanding towards the existence state of Cu species was obtained by EPR, shown in Fig. 6. Only  $\text{Cu}^{2+}$  species with  $d^9$  electronic structure gave EPR signal, whereas other Cu valence were EPR silent [34]. The signal A was a typical peak of monomeric  $\text{Cu}^{2+}$  in octahedral sites with a tetragonal distortion [35]. The signal K at 3000 G was assigned to dimeric  $\text{Cu}^{2+}$  generated from the coupled unpaired electrons of two  $\text{Cu}^{2+}$  [36]. It's quite clear that both the  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  and  $\text{Cu}_{0.030}/\text{Nb}_1\text{Ce}_3$  had much more monomeric  $\text{Cu}^{2+}$  species. Besides, the distinct K signal of  $\text{Cu}_{0.030}/\text{Nb}_1\text{Ce}_3$  provided evidence for the excessive Cu addition induced agglomeration of Cu species.



**Fig. 6.** EPR spectra of  $\text{Nb}_1\text{Ce}_3$  and  $\text{Cu}/\text{Nb}_1\text{Ce}_3$ .

To further confirm the surface state and content of Cu species, *in situ* DRIFTS was

performed for NO chemisorption. The interaction of NO with  $\text{Cu}^{2+}$  and  $\text{Cu}^+$  would give specific bands, so it's a reasonable semi-quantitative means to determine the  $\text{Cu}^{2+}/(\text{Cu}^{2+}+\text{Cu}^+)$  ratio.  $\text{Cu}/\text{Nb}_1\text{Ce}_3$  catalysts were exposed to flowing  $\text{NO}/\text{N}_2$  until saturated at room temperature and the result was shown in Fig. 7. The band detected at *ca.*  $1905\text{ cm}^{-1}$  was ascribed to the  $\text{Cu}^{2+}\text{-NO}$  and *ca.*  $1847\text{ cm}^{-1}$  was ascribed to the  $\text{Cu}^+\text{-NO}$  [37]. Later, the peak area was integrated respectively and the  $\text{Cu}^{2+}/(\text{Cu}^{2+}+\text{Cu}^+)$  ratio was further calculated as placed in Table 3. Obviously, the  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  catalyst had the highest  $\text{Cu}^{2+}/(\text{Cu}^{2+}+\text{Cu}^+)$  ratio.



**Fig. 7.** *in situ* DRIFTS spectra of  $\text{Cu}/\text{Nb}_1\text{Ce}_3$  saturated in NO at room temperature.

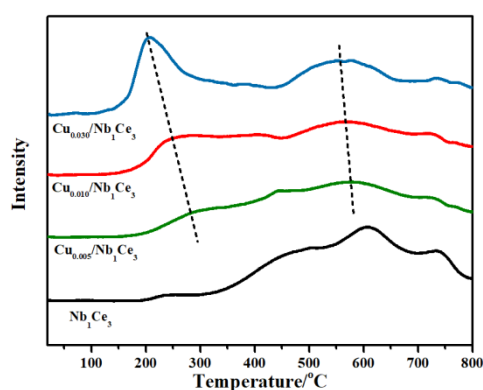
**Table 3.** Surface  $\text{Cu}^{2+}/(\text{Cu}^{2+}+\text{Cu}^+)$  ratio of  $\text{Cu}/\text{Nb}_1\text{Ce}_3$  derived from NO-*in situ* DRIFTS

| Samples                                    | $\text{Cu}^{2+}/(\text{Cu}^{2+}+\text{Cu}^+)/\%$ |
|--|--|
| $\text{Cu}_{0.005}/\text{Nb}_1\text{Ce}_3$ | 55.5   |
| $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$ | 58.9   |
| $\text{Cu}_{0.030}/\text{Nb}_1\text{Ce}_3$ | 53.0   |

### 3.4 Clarification of Cu promoting mechanism

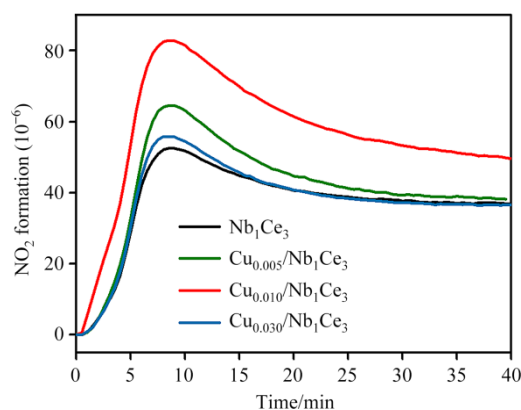
$\text{H}_2$ -TPR was performed to reveal the promotional effect of Cu modification on the

redox properties of  $\text{Nb}_1\text{Ce}_3$  and the results were reported in Fig. 8. A broad reduction peak occurred on the  $\text{Nb}_1\text{Ce}_3$  surface from 400 °C to 750 °C, which could be associated with the reduction of the surface capping oxygen [20, 38]. For  $\text{Cu}/\text{Nb}_1\text{Ce}_3$ , a new reduction peak centered at about 200-300 °C occurred, representing that the  $\text{Cu}^{2+}$  was reduced to  $\text{Cu}^+$  and then  $\text{Cu}^+$  was converted to metallic Cu [39], suggesting the significant enhancement of redox ability in low-temperature range. It's observed that the reduction peak of Cu species shifted to lower temperature with the Cu addition increased, proving that the gradually enhanced redox properties of  $\text{Cu}/\text{Nb}_1\text{Ce}_3$ . Besides, the peak of surface capping oxygen moved to the lower temperature due to the interaction among Cu, Nb and Ce species. Nevertheless, excessive redox capacity would lead to the occurrence of  $\text{NH}_3$ -SCR side reactions at high temperature, resulting in diminishment for NO removal [31]. Therefore, it could be inferred that the strong redox property generated from agglomerated Cu species on  $\text{Cu}_{0.030}/\text{Nb}_1\text{Ce}_3$  made the weakened denitration efficiency at high temperature [40]. However, an appropriate redox property generated from highly dispersed isolated Cu species enabled  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  to acquire a promoted SCR performance over a wide temperature.



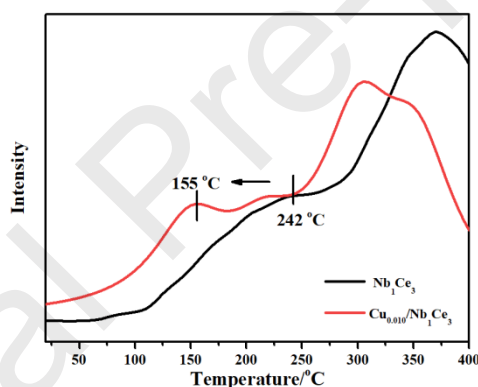
**Fig. 8.** H<sub>2</sub>-TPR profiles of Nb<sub>1</sub>Ce<sub>3</sub> and Cu/Nb<sub>1</sub>Ce<sub>3</sub>.

Cu modification efficaciously enhanced the denitration efficiency of Nb<sub>1</sub>Ce<sub>3</sub> and the Cu<sub>0.010</sub>/Nb<sub>1</sub>Ce<sub>3</sub> catalyst with the most Cu<sup>2+</sup> exhibited the best activity. Hence, the effect of Cu<sup>2+</sup> on the denitrification reaction was worthy of further investigation. Recently studies showed that for Cu loaded zeolite catalysts, Cu<sup>2+</sup> would induce the Cu<sup>2+</sup>-NO<sub>3</sub><sup>-</sup>+NO(g)→Cu<sup>2+</sup>-NO<sub>2</sub><sup>-</sup>+NO<sub>2</sub>(g) reaction, then the generated NO<sub>2</sub> promoted fast-SCR reaction and improved low-temperature NH<sub>3</sub>-SCR reaction [37]. Herein, an experiment of NO reduction of surface nitrates was conducted to explore the promotional effect of Cu<sup>2+</sup>. NO+O<sub>2</sub>/N<sub>2</sub> were pre-adsorbed on Cu/Nb<sub>1</sub>Ce<sub>3</sub> catalysts until saturated at 150 °C so that nitrates were formed on the surface. Then the catalyst was exposed to the NO after purged in inert atmosphere and the NO<sub>2</sub> formation was recorded in Fig. 9. It could be clearly seen that the NO<sub>2</sub> formation increased after Cu modification and the promoting effect was volcanic with the increase of Cu<sup>2+</sup> content. Cu<sub>0.010</sub>/Nb<sub>1</sub>Ce<sub>3</sub> catalyst with the most Cu<sup>2+</sup> activated and reduced more surface nitrates to generate the most NO<sub>2</sub>. This helped to reveal the promotional mechanism of Cu<sup>2+</sup> on the low-temperature deNO<sub>x</sub> performance of Cu<sub>0.010</sub>/Nb<sub>1</sub>Ce<sub>3</sub> catalyst.



**Fig. 9.** NO<sub>2</sub> formation of Nb<sub>1</sub>Ce<sub>3</sub> and Cu/Nb<sub>1</sub>Ce<sub>3</sub>.

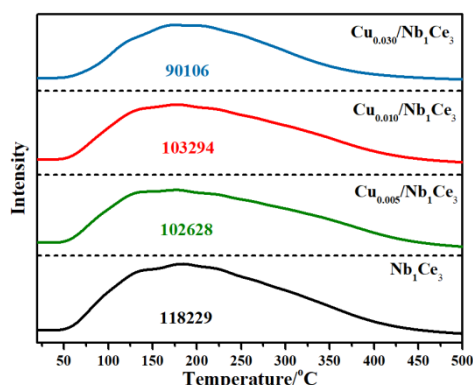
It's known that a key step in the catalytic reaction was the desorption process of the product, so the NO<sub>2</sub> adsorption capacity was further evaluated by NO<sub>2</sub>-TPD. As shown in Fig. 10, after Cu modification, the desorption temperature of ad-NO<sub>2</sub><sup>-</sup> [41] over Cu<sub>0.010</sub>/Nb<sub>1</sub>Ce<sub>3</sub> catalyst decreased significantly, from 242 °C to 155 °C. The easier the NO<sub>2</sub> desorbed, the faster the SCR reaction carried out. Besides, the desorption peak at high temperature was related to the strongly adsorbed nitrates [42], their easy desorption would also be beneficial to more sites available for NH<sub>3</sub> adsorption. And the NO elimination at high temperature was further facilitated.

**Fig. 10.** NO<sub>2</sub>-TPD profiles of Nb<sub>1</sub>Ce<sub>3</sub> and Cu<sub>0.010</sub>/Nb<sub>1</sub>Ce<sub>3</sub>.

### 3.5 Adsorption properties

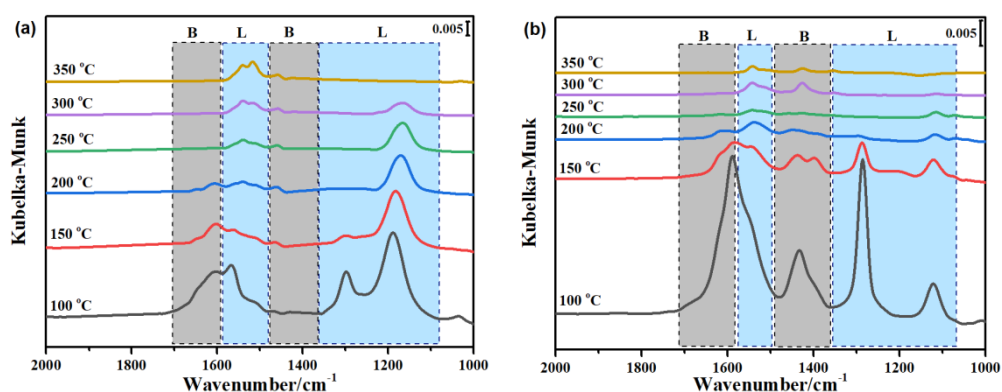
Surface acidity was another decisive factor for the NH<sub>3</sub>-SCR reaction. Therefore, NH<sub>3</sub>-TPD was performed to determine the amount of acid sites and acid strength. As shown in Fig. 11, all samples showed a broad desorption peak in the range of 50–450 °C and the profiles changed little after Cu modification. The integrated peak areas manifested that the Cu<sub>0.010</sub>/Nb<sub>1</sub>Ce<sub>3</sub> catalyst showed the highest NH<sub>3</sub> adsorption capacity

among all Cu modified NbCe catalysts. And the decrease in total acid sites probably caused by the coverage of agglomerated Cu species.



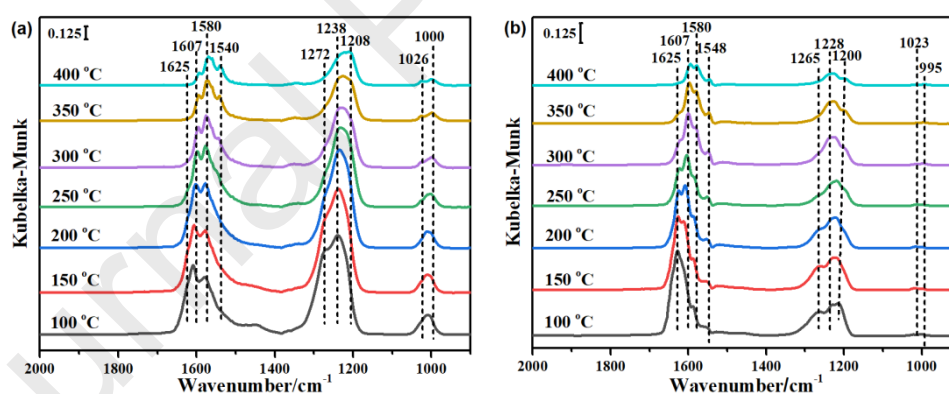
**Fig. 11.** NH<sub>3</sub>-TPD profiles of Nb<sub>1</sub>Ce<sub>3</sub> and Cu/Nb<sub>1</sub>Ce<sub>3</sub>.

To further distinguish the specific adsorption sites of NH<sub>3</sub>, Fig. 12 illustrated the results of *in situ* DRIFTS spectra of NH<sub>3</sub>-TPD over (a) Nb<sub>1</sub>Ce<sub>3</sub> and (b) Cu<sub>0.010</sub>/Nb<sub>1</sub>Ce<sub>3</sub>. For Nb<sub>1</sub>Ce<sub>3</sub> catalyst, the bands centered at *ca.* 1604, 1568, 1297 and 1190 cm<sup>-1</sup> could be ascribed to bending vibrations of N-H bonds in NH<sub>3</sub> coordinately linked to Lewis acid sites [43, 44], while the weak band at *ca.* 1612 and 1458 cm<sup>-1</sup> was assigned to bending vibrations of ionic NH<sub>4</sub><sup>+</sup> on the Brønsted acid sites [6, 38]. For Cu<sub>0.010</sub>/Nb<sub>1</sub>Ce<sub>3</sub> catalyst, bands ascribed to Brønsted acid sites were distinctly enhanced on account of the Cu modification. This should be caused by the appearance of NbO<sub>4</sub> species due to the strong interaction between Nb and Cu species as Raman spectra indicated [26].



**Fig. 12.** *in situ* DRIFTS spectra of (a)  $\text{Nb}_1\text{Ce}_3$  and (b)  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  saturated in  $\text{NH}_3$  at 100 °C and then purged by  $\text{N}_2$  from 100 °C to 350 °C.

The  $\text{NO}+\text{O}_2$ -TPD over (a)  $\text{Nb}_1\text{Ce}_3$  and (b)  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  were also carried out through *in situ* DRIFTS. As depicted in Fig. 13, the bands of bridging bidentate nitrates (1625, 1607 and 1208  $\text{cm}^{-1}$ ), chelating bidentate nitrates (1580, 1540 and 1238  $\text{cm}^{-1}$ ), monodentate nitrates (1548  $\text{cm}^{-1}$ ) and *cis*-( $\text{N}_2\text{O}_2$ )<sup>2-</sup> (1026 and 1000  $\text{cm}^{-1}$ ) were observed [38, 44]. It could be found that after Cu modification, the ad-nitrate species were almost unchanged, but the thermal stability of the nitrates on  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  catalyst was remarkably reduced. This manifested that the nitrates adsorbed on  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  was more easily activated to participate in the following reaction, and thus the adsorption and activation sites would be more accessible for  $\text{NH}_3$  at higher temperature.

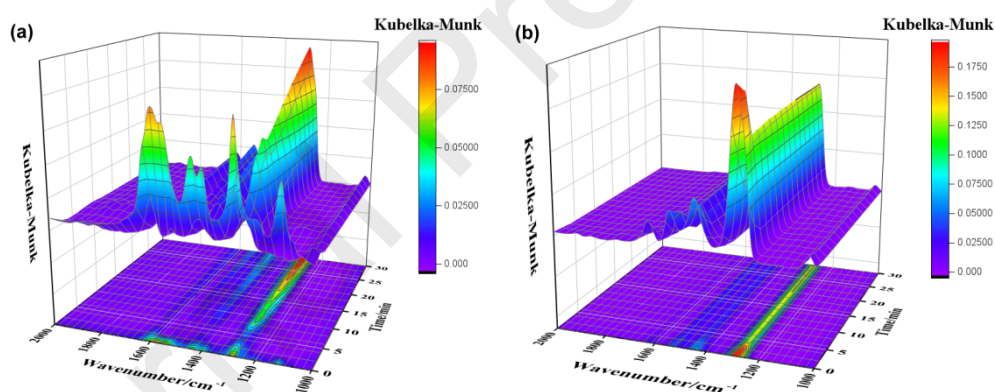


**Fig. 13.** *in situ* DRIFTS spectra of (a)  $\text{Nb}_1\text{Ce}_3$  and (b)  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  saturated in  $\text{NO}+\text{O}_2$  at 100 °C and then purged by  $\text{N}_2$  from 100 °C to 350 °C.

### 3.6 Reaction mechanism

To determine the reactivity of  $\text{NH}_3$  on  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$ , *in situ* DRIFTS was conducted at 150 °C with a flowing stream of  $\text{NO}+\text{O}_2$  reacting with preadsorbed  $\text{NH}_3$

(Fig. 14(a)). Upon introducing  $\text{NO}+\text{O}_2$  to catalyst preadsorbed with  $\text{NH}_3$ , the bands assigned to  $\text{NH}_3$  on Lewis and Brønsted acid sites consumed rapidly, suggesting that the SCR reaction on  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  could proceed through the Eley-Rideal (E – R) mechanism efficiently. In turn, the reactivity of adsorbed nitrates on  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  was also studied with  $\text{NH}_3$  reacting with preadsorbed  $\text{NO}+\text{O}_2$  (Fig. 14(b)). The bands assigned to nitrate species (bridged nitrate at  $1621\text{ cm}^{-1}$ , monodentate nitrates at  $1539$ ,  $1442$  and  $1290\text{ cm}^{-1}$ ) were displayed. Afterwards,  $\text{NH}_3$  were introduced, the intensity of bridge nitrate was weakened and the monodentate nitrate was reconstructed, indicating that the adsorbed nitrates could not react with  $\text{NH}_3$ . Therefore, the SCR reaction followed Eley-Rideal (E-R) mechanism over  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  at  $150\text{ }^\circ\text{C}$ .



**Fig. 14.** *in situ* DRIFTS spectra of  $\text{NH}_3$ -SCR reaction on  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  at  $150\text{ }^\circ\text{C}$ .

#### 4. Conclusions

A series of Cu modified NbCe catalysts were prepared and applied in the  $\text{NH}_3$ -SCR process. The ternary oxides catalysts presented promising NO conversion in a broad reaction temperature window compared with binary oxides NbCe catalyst. Among all the synthesized catalysts, the  $\text{Cu}_{0.010}/\text{Nb}_1\text{Ce}_3$  catalyst showed the optimal

NO abatement efficiency from 200 to 400 °C, together with near 100% N<sub>2</sub> selectivity. Besides, H<sub>2</sub>O or/and SO<sub>2</sub> had a slightly reversible inhibition influence on the catalytic activity. Several conclusions can be drawn through an array of characterization: The proper amount of Cu modification made the active Cu species highly dispersed in the form of monomeric or dimeric Cu<sup>2+</sup> on Cu<sub>0.010</sub>/Nb<sub>1</sub>Ce<sub>3</sub>, facilitating the less stable nitrates, the formation/desorption of NO<sub>2</sub> and the fast-SCR reaction. Meanwhile, the interaction between Cu and Nb species induced the generation of highly active NbO<sub>4</sub> species and more surface acid sites, therefore the adsorption and activation of NH<sub>3</sub> were enhanced, which were all responsible for the excellent NH<sub>3</sub>-SCR activity. Nevertheless, excessive Cu addition would lead to the appearance of agglomerated Cu species, resulting in undesirable NH<sub>3</sub> oxidation and eventually impaired deNO<sub>x</sub> efficiency at high temperature.

### **Declaration of Competing Interest**

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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### **Declaration of Competing Interest**

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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